

# Enhanced precipitation variability decreases grass- and increases shrub-productivity

Laureano A. Gherardi<sup>a,1</sup> and Osvaldo E. Sala<sup>a,b</sup>

<sup>a</sup>School of Life Sciences, Arizona State University, Tempe, AZ 85287-4501; and <sup>b</sup>School of Sustainability, Arizona State University, Tempe, AZ, 85287-4501

Edited by William H. Schlesinger, Cary Institute of Ecosystem Studies, Millbrook, NY, and approved August 25, 2015 (received for review April 1, 2015)

Although projections of precipitation change indicate increases in variability, most studies of impacts of climate change on ecosystems focused on effects of changes in amount of precipitation, overlooking precipitation variability effects, especially at the interannual scale. Here, we present results from a 6-y field experiment, where we applied sequences of wet and dry years, increasing interannual precipitation coefficient of variation while maintaining a precipitation amount constant. Increased precipitation variability significantly reduced ecosystem primary production. Dominant plant-functional types showed opposite responses: perennial-grass productivity decreased by 81%, whereas shrub productivity increased by 67%. This pattern was explained by different nonlinear responses to precipitation. Grass productivity presented a saturating response to precipitation where dry years had a larger negative effect than the positive effects of wet years. In contrast, shrubs showed an increasing response to precipitation that resulted in an increase in average productivity with increasing precipitation variability. In addition, the effects of precipitation variation increased through time. We argue that the differential responses of grasses and shrubs to precipitation variability and the amplification of this phenomenon through time result from contrasting root distributions of grasses and shrubs and competitive interactions among plant types, confirmed by structural equation analysis. Under drought conditions, grasses reduce their abundance and their ability to absorb water that then is transferred to deep soil layers that are exclusively explored by shrubs. Our work addresses an understudied dimension of climate change that might lead to widespread shrub encroachment reducing the provisioning of ecosystem services to society.

ANPP | precipitation | interannual variability | plant-functional types | nonlinear response

Climate-change simulations project increases in precipitation variability as a result of global warming (1–3). The frequency of large precipitation events is expected to increase (3, 4), even in regions where precipitation will decrease (5). Similarly, the occurrence of wet days will decrease, resulting in a highly variable climate with enhanced probabilities of drought and heavy rainfall (5). Precipitation change will occur at intra-, interannual, and decadal scales. Mechanisms explaining such changes differ among temporal scales. At short-time scales, high precipitation variation results from the increased water-holding capacity of a warmer atmosphere that yields large rainfall events interspaced with droughts (6). At the interannual and decadal scales, climate change results in enhanced precipitation variability resulting from changes in atmospheric circulation that affect multiyear rainfall patterns (7).

Although precipitation variability changes are part of the public narrative (8) and motivated a special Intergovernmental Panel on Climate Change (IPCC) report on extreme events (9), our understanding of the effect of climate variability on the carbon cycle in grasslands is still weak (10). Aboveground net primary production (ANPP) is the main carbon fixation pathway, and even though its responses to changes in the amount of precipitation have been well studied, knowledge about the effect of precipitation variability on ANPP is rather poor, especially at the interannual

to decadal time scales. A few short-term experiments focused on the effect of intra-annual precipitation variance and reported contrasting results, with null or positive effects in arid systems and negative effects in mesic systems (11–14). A modeling exercise found that increased interannual precipitation variability and temperature in the Tibetan Plateau led to productivity reduction in grasslands (15). Therefore, long-term manipulative experiments are needed to understand ecosystem responses to changes in resource amount and variability (16) at multiyear time scales.

Here, we aimed at assessing the effect interannual variability of precipitation on ecosystem functioning. Two hypotheses guided our work. (i) Increased interannual precipitation variation may increase, decrease, or have null effect on ANPP, depending on the shape of the productivity response to precipitation. Increased interannual precipitation variability implies sequences of relatively extreme dry and wet years. If the ANPP response to annual precipitation is linear, increased precipitation variance will result in a null effect on mean multiyear ANPP because the decline caused by dry years is compensated by ANPP increases in wet years. On the other hand, nonlinear responses result in either positive or negative effects of precipitation variation on ANPP (17). For example, an increasing ANPP response to annual precipitation leads to positive effects of precipitation variability because the ANPP decline caused by dry years is overcompensated by the nonlinear ANPP increase resulting from wet years. A decreasing response results in negative response to precipitation variability because the positive effect of wet years does not compensate dry-year ANPP decreases (18). (ii) The effect of enhanced interannual precipitation variation increases through time. The duration of increased precipitation variability periods may lead to hierarchical ecosystem responses (19). We expect that physiological responses

## Significance

Although increased climatic variability resulting from climate change has been accepted by the scientific community and forms part of the public narrative, studies of the effect of climatic variability on ecosystems have received much less attention than effects of changes in the mean state of climate. Here, we report on a field experiment where we experimentally increased interannual variability of precipitation while maintaining average precipitation constant. Our results indicated that total productivity and that of grasses declined in response to increased precipitation variability, although shrubs benefited, suggesting a potential shift from grassland to shrubland in the future, with large consequences for the supply of ecosystem services. In addition, the effect of variability increased through time.

Author contributions: L.A.G. and O.E.S. designed research, performed research, analyzed data, and wrote the paper.

The authors declare no conflict of interest.

This article is a PNAS Direct Submission.

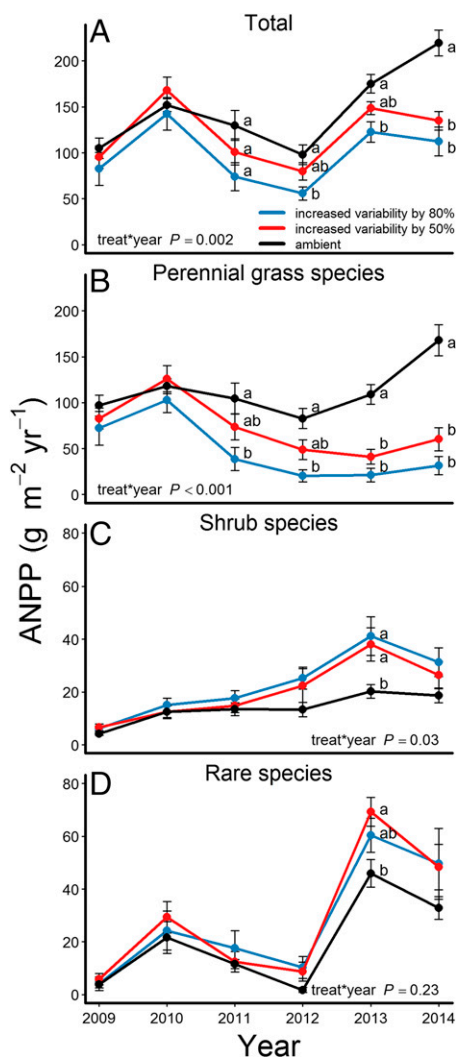
<sup>1</sup>To whom correspondence should be addressed. Email: Lau@ASU.edu.

This article contains supporting information online at [www.pnas.org/lookup/suppl/doi:10.1073/pnas.1506433112/-DCSupplemental](http://www.pnas.org/lookup/suppl/doi:10.1073/pnas.1506433112/-DCSupplemental).



The ANPP difference between the highest variability treatments and the ambient precipitation treatment increased from  $16 \text{ g}\cdot\text{m}^{-2}\cdot\text{yr}^{-1}$  at the beginning of the experiment to  $107 \text{ g}\cdot\text{m}^{-2}\cdot\text{yr}^{-1}$  at the end of 6 y of treatment. We hypothesize that the response of ANPP to increased interannual precipitation variability is amplified through time because of the gradual engagement of different





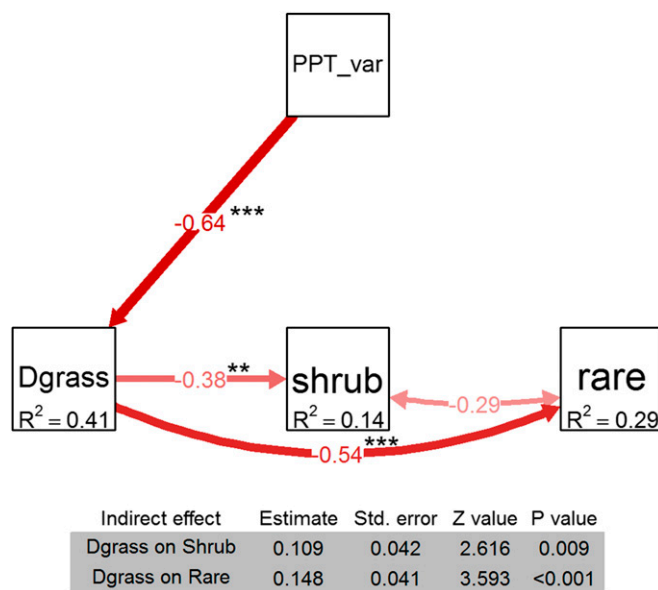
**Fig. 4.** Effect of precipitation variability on ANPP through time for: (A) total ANPP, (B) dominant-grass species ANPP, (C) shrub-species ANPP, and (D) rare-species ANPP. Points indicate mean values ( $\pm$ SE,  $n = 10$ ) for each treatment; and different colors indicate different treatments. Different letters indicate significant differences among treatments within year. No letter means nonsignificant difference.

the third year until the end of the experiment (Fig. 4B). In contrast, shrub and rare species responses to increased precipitation-variability treatments were different from control only at the end of the experiment (Fig. 4C and D and *SI Appendix, Statistical Analysis Description and Summary Output*).

We argue that the opposite responses of shrubs and grasses to precipitation variability result from contrasting root distributions and competitive interactions. Shrubs have deep roots (25) and use water stored in deep soil layers (26), whereas grasses have relatively shallow roots and use soil water located in upper layers of the soil (27). Therefore, changes in the location of available resources may determine the competitive balance between the two plant types (25). Increased precipitation variability has been shown to shift the soil profile downward (28), potentially explaining the positive effect on deep-rooted shrubs. In addition to the location of water sources, these contrasting rooting patterns determine the volume of soil explored by each plant type. Shrubs explore a relatively large volume of soil where water from wet years can be stored, whereas grasses not only explore a smaller volume of soil, but excess water percolating from top layers during wet years recharges the portion of the soil explored by shrubs. This mechanism

explains why shrubs are benefited by increased variability but grasses are negatively affected. Low-to-modal precipitation years fill the top layer of the soil and have a positive effect on grass ANPP. Extremely wet years overcome the soil water-holding capacity of the top layer, explaining the plateau in the grass ANPP response, and percolate deeper into the soil profile, causing the increase in shrub ANPP (Fig. 3).

The mechanism explaining the observed amplification of the ecosystem response to enhanced precipitation variability through time is associated with biotic interactions among grasses, shrubs, and the rare-species. We ran a structural-equation model to specifically test the indirect effect of dominant-grass ANPP on the productivity of shrub and rare species (Fig. 5). The model showed positive indirect effects of precipitation variation on shrub and rare species through its negative effect on perennial-grass ANPP (*SI Appendix, Statistical Analysis Description and Summary Output*). This result supports the idea that shrubs and rare species benefit from the diminished dominance of perennial grasses under high precipitation variation and explain the delayed responses of shrub and rare species behind dominant grasses. During wet years following dry years, grasses do not hold enough structures to efficiently use available resources in the upper layers of the soil (24). Therefore, underused resources may percolate deep in the soil profile, increasing the pool of available resources in deep soil layers (29), enhancing shrub ANPP. The amplification results from each dry cycle reducing grass ability to capture resources that are transferred to shrubs. The longer the duration of the enhanced precipitation-variability conditions,



**Fig. 5.** Effect of precipitation coefficient of variation on 6-y mean dominant grass ANPP (Dgrass) and its indirect effects on 6-y-mean shrub and rare-species ANPP. The model includes the direct effect of precipitation coefficient of variation on dominant grasses and indirect effects through dominant grass ANPP on shrubs and rare species ANPP. Path coefficients are standardized by the mean so they are comparable to each other. Single-headed arrows mean direct negative effects. Double-headed arrows indicate noncausal correlation. Indirect effects result from the multiplication of two consecutive direct effects. (Inset) Table shows coefficients for indirect effects including estimates, SE, z scores, and P values. The model is well supported by our data ( $\chi^2 = 2.47$ ,  $df = 2$ ,  $P = 0.29$ ). Other goodness-of-fit measures also support this model (standardized root mean square residual = 0.04, root mean square error = 0.07, comparative fit index = 0.974, Tucker–Lewis index = 0.973). Significance codes mean: ns, not significant, \* $P < 0.05$ , \*\* $P < 0.01$ , and \*\*\* $P < 0.001$ . For detailed description of analysis and output, see *SI Appendix, Statistical Analysis Description and Summary Output*.

the smaller the capacity of grasses to absorb water and the larger the transfer to deep soil layers where shrub have exclusive access. Rare species may benefit from reduced direct competition because of reduced abundance of the dominant grass.

The magnitude of the reduction in primary production because of increased precipitation variability is considerably large. ANPP in high precipitation-variability treatments decreased 49%; where grass ANPP decreased 81%, shrub ANPP increased 67%, and rare-species ANPP increased 50% compared with control ANPP in the sixth year of the experiment. We demonstrated here that ecosystem response to precipitation variability increased through time along the duration of our experiment. The effect of further ecosystem exposure to high interannual-precipitation variance is still uncertain. It may lead to a plateau if the precipitation variability effect is constrained by the water-holding capacity of deep soil layers. Alternatively, the positive feedback between the demise of grasses and the positive response of shrubs may lead the ecosystem into a nonlinear trajectory. If the decline of dominant grass species continues, the ecosystem may transition into a new state of shrub dominance with lower productivity (30, 31). Arid and semiarid ecosystems occupy a large fraction of the global terrestrial land so these drastic impacts of enhanced climate variability may have global consequences (32, 33).

## Methods

**Site Description.** The experiment was carried out at the Jornada Basin Long-Term Ecological Research site (32.5°N, 106.8°W, 1,188 m above sea level) in New Mexico, United States. Long-term mean growing-season precipitation is 105 mm, with a coefficient of variation for a 6-y time window of 38% that ranges from 18% to 67%. The desert grassland under study is dominated by *Bouteloua eriopoda* (Torr.), with the presence of *Prosopis glandulosa* (Torr.). Soils are coarse-textured and present a petrocalcic horizon at depths ranging from 64 to 76 cm (34).

**Experimental Design.** We increased interannual precipitation variability by alternating rainfall interception and irrigation for 6 y, switching treatments every spring before the growing season started. Our treatments were designed to be relative to ambient precipitation, so we kept the number and timing of rainfall pulses under natural conditions and constant across treatments. The reasoning behind our design was to isolate the effect of precipitation variance from the effects of pulse number and timing studied elsewhere. With this design, we also kept constant among plots all other climatic factors. In fact, the product of natural rainfall and our treatments resulted in precipitation-variation treatments ranging from 48% to 110%; mean precipitation stayed almost the same among treatments, ranging from 126.7 mm to 128.5 mm. We used the automated rainfall manipulation system, ARMS (20), which consisted of rainout shelters (35) that collected either 50% or 80% of the incoming rainfall from exclusion plots and diverted it by means of a solar-powered pump to irrigation plots; control plots received ambient precipitation throughout the duration of the experiment. See refs. 20 and 35 for rainfall manipulation details.

Our experiment consisted of 10 replicates of 5 levels of precipitation interannual coefficient of variation, totaling 50 plots of 2.5 m by 2.5 m that were trenched down to 60 cm or to petrocalcic layer, and lined with 6-mL PVC film to prevent lateral movement of water and roots in or out of the plots. We ensured that all plots had the same starting conditions by picking 80 plots and keeping the 50 most similar in terms of plant-type cover and assigning treatments to each plot randomly.

## Response Variables.

**ANPP.** Plant species were classified into plant types on the basis of their contribution to total productivity. Annual grasses, forbs, and subshrubs form the rare-species group and have low biomass, make small contribution to ANPP, and show episodic growth and reproduction. Dominant grasses represent the main component of the ecosystem productivity, are shallow-rooted and short-lived perennials. The shrub species present in our study is the second dominant, long-lived perennial species, and deep-rooted.

To avoid clipping effects in our multiyear experiment, we estimated ANPP using a nondestructive method that uses plant-species cover and shrub volume (36) as proxies for ANPP. We estimated herbaceous-species plant cover to a 1-cm precision on three 2.5-m permanent cover lines per plot; shrub volume was estimated by measuring two perpendicular diameters and height. ANPP was derived using allometric equations for each plant functional type developed on site. In the case of rare species, we did allometric calibrations for annual grasses, forbs, and subshrubs and estimated cover and harvested nine 20-cm by 40-cm quadrats of nine species (three annual grasses, three forbs, and three subshrubs), totaling 81 quadrats at peak biomass. Twenty 0.2-m by 1-m quadrats were double-sampled for perennial grasses. Biomass-cover regressions were run and slope estimates were used to transform species cover into ANPP. We also developed a shrub allometric equation harvesting one-quarter of 24 shrubs. We sorted for current-year biomass, dried and weighed the samples, and fit a regression model of measured ANPP values against shrub volume (Fig. S3). We are confident of our ANPP estimation method because it matches long-term measurements done with a different method at a similar site within the Jornada Long-Term Ecological Research. For example, in 2009 the mean ANPP for our control plots was 104.8 g·m<sup>-2</sup>·yr<sup>-1</sup> and ANPP for the International Biological Programme Enclosure (IBPE) site (similar vegetation structure to our site) was 103.2 g·m<sup>-2</sup>·yr<sup>-1</sup>.

**Soil moisture.** We measured soil moisture in the top 30 cm of the soil profile every 30 min in four replicates of each treatment using Campbell Scientific CS625 probes and the data were logged onto CR200X data loggers during 3 y of the experiment (Fig. S1).

**Statistical Analyses.** We performed all analyses and created all figures using R v3.0.2 (37). For detailed description of statistical analyses and correspondent output, see *SI Appendix, Statistical Analysis Description and Summary Output*.

**ACKNOWLEDGMENTS.** We thank L. G. Reichmann, L. Yahdjian, N. M. Sala, and P. Flombaum for earlier discussion of ideas and input; J. Angel III, J. P. Midez, P. B. R. McKenna, K. Duffy, J. Haussler, and B. O. Brothers for field assistance; G. A. Gil for guidance; the Jornada Basin Long-Term Ecological Research leaders and personnel for all their support; and Dan Childers, Nancy Grimm, Sharon Hall, and Jingle Wu for their contributions. This research was supported by National Science Foundation Grants DEB-1235828 and DEB 09-17668.

- Fischer EM, Beyerle U, Knutti R (2013) Robust spatially aggregated projections of climate extremes. *Nat Clim Chang* 3(12):1033–1038.
- Wetherald R (2010) Changes of time mean state and variability of hydrology in response to a doubling and quadrupling of CO<sub>2</sub>. *Clim Change* 102(3–4):651–670.
- Räisänen J (2002) CO<sub>2</sub>-induced changes in interannual temperature and precipitation variability in 19 CMIP2 experiments. *J Clim* 15(17):2395–2411.
- IPCC (2013) Climate Change 2013: The physical science basis. Contribution of working group I to *The Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, eds Stocker TF, et al. (Cambridge Univ Press, Cambridge, UK, and New York), pp 1–1535.
- Kharin VV, Zwiers FW, Zhang XB, Hegerl GC (2007) Changes in temperature and precipitation extremes in the IPCC ensemble of global coupled model simulations. *J Clim* 20(8):1419–1444.
- Allen MR, Ingram WJ (2002) Constraints on future changes in climate and the hydrologic cycle. *Nature* 419(6903):224–232.
- Trenberth KE, Dai A, Rasmussen RM, Parsons DB (2003) The changing character of precipitation. *Bull Am Meteorol Soc* 84(9):1205–1218.
- Leiserowitz A, Maibach E, Roser-Renouf C, Hmielowski J (2012) *Extreme Weather, Climate & Preparedness in the American Mind* (Yale Univ and George Mason Univ, New Haven, CT).
- IPCC (2012) Managing the risks of extreme events disasters to advance climate change adaptation. *A Special Report of Working Groups I and II of the Intergovernmental Panel on Climate Change*, eds Field CB, et al. (Cambridge Univ Press, Cambridge, UK, and New York), pp 1–582.
- Reichstein M, et al. (2013) Climate extremes and the carbon cycle. *Nature* 500(7462):287–295.
- Knapp AK, et al. (2002) Rainfall variability, carbon cycling, and plant species diversity in a mesic grassland. *Science* 298(5601):2202–2205.
- Thomey ML, et al. (2011) Effect of precipitation variability on net primary production and soil respiration in a Chihuahuan Desert grassland. *Glob Change Biol* 17(4):1505–1515.
- Heisler-White JL, Blair JM, Kelly EF, Harmoney K, Knapp AK (2009) Contingent productivity responses to more extreme rainfall regimes across a grassland biome. *Glob Change Biol* 15(12):2894–2904.
- Fay PA, Kaufman DM, Nippert JB, Carlisle JD, Harper CW (2008) Changes in grassland ecosystem function due to extreme rainfall events: Implications for responses to climate change. *Glob Change Biol* 14(7):1600–1608.
- Ye J-S, Reynolds J, Sun G-J, Li F-M (2013) Impacts of increased variability in precipitation and air temperature on net primary productivity of the Tibetan Plateau: A modeling analysis. *Clim Change* 119(2):321–332.

16. Smith MD, et al. (2015) Global environmental change and the nature of aboveground net primary productivity responses: Insights from long-term experiments. *Oecologia* 177(4):935–947.
17. Hsu JS, Powell J, Adler PB (2012) Sensitivity of mean annual primary production to precipitation. *Glob Change Biol* 18(7):2246–2255.
18. Hsu JS, Adler PB (2014) Anticipating changes in variability of grassland production due to increases in interannual precipitation variability. *Ecosphere* 5(5):art58.
19. Collins SL, et al. (2014) A multiscale, hierarchical model of pulse dynamics in arid-land ecosystems. *Annu Rev Ecol Syst* 45(1):397–419.
20. Gherardi LA, Sala OE (2013) Automated rainfall manipulation system: A reliable and inexpensive tool for ecologists. *Ecosphere* 4(2):art18.
21. Reichmann LG, Sala OE, Peters DPC (2013) Precipitation legacies in desert grassland primary production occur through previous-year tiller density. *Ecology* 94(2):435–443.
22. Sala OE, Gherardi LA, Reichmann LG, Jobbágy E, Peters D (2012) Legacies of precipitation fluctuations on primary production: Theory and data synthesis. *Phil Trans R Soc Lond B Biol Sci* 367(1605):3135–3144.
23. Throop HL, Reichmann LG, Sala OE, Archer SR (2012) Response of dominant grass and shrub species to water manipulation: An ecophysiological basis for shrub invasion in a Chihuahuan Desert grassland. *Oecologia* 169(2):373–383.
24. Reichmann LG, Sala OE (2014) Differential sensitivities of grassland structural components to changes in precipitation mediate productivity response in a desert ecosystem. *Funct Ecol* 28(5):1292–1298.
25. Weltzin JF, McPherson GR (2000) Implications of precipitation redistribution for shifts in temperate savanna ecotones. *Ecology* 81(7):1902–1913.
26. Sala OE, Lauenroth WK, Parton WJ (1992) Long term soil water dynamics in the shortgrass steppe. *Ecology* 73(4):1175–1181.
27. Jackson RB, et al. (1996) A global analysis of root distributions for terrestrial biomes. *Oecologia* 108(3):389–411.
28. Sala O, Gherardi L, Peters DC (2015) Enhanced precipitation variability effects on water losses and ecosystem functioning: Differential response of arid and mesic regions. *Clim Change* 131(2):213–227.
29. Sala OE, Golluscio RA, Lauenroth WK, Soriano A (1989) Resource partitioning between shrubs and grasses in the Patagonian Steppe. *Oecologia* 81(4):501–505.
30. Archer S (2010) Rangeland conservation and shrub encroachment: New perspectives on an old problem. *Wild Rangelands: Conserving Wildlife While Maintaining Livestock in Semi-Arid Ecosystems*, eds du Toit JT, Kock R, Deutsch JC (John Wiley & Sons, Chichester, UK), pp 53–97.
31. Anadón JD, Sala OE, Turner BL, 2nd, Bennett EM (2014) Effect of woody-plant encroachment on livestock production in North and South America. *Proc Natl Acad Sci USA* 111(35):12948–12953.
32. Poulter B, et al. (2014) Contribution of semi-arid ecosystems to interannual variability of the global carbon cycle. *Nature* 509(7502):600–603.
33. Ahlström A, et al. (2015) Carbon cycle. The dominant role of semi-arid ecosystems in the trend and variability of the land CO<sub>2</sub> sink. *Science* 348(6237):895–899.
34. Havstad KM, Schlesinger WH (2006) Introduction. *Structure and Function of a Chihuahuan Desert Ecosystem*, eds Havstad KM, Schlesinger WH, Huenneke LF (Oxford Univ Press, Oxford), pp 3–15.
35. Yahdjian L, Sala OE (2002) A rainout shelter design for intercepting different amounts of rainfall. *Oecologia* 133(2):95–101.
36. Flombaum P, Sala OE (2007) A non-destructive and rapid method to estimate biomass and aboveground net primary production in arid environments. *J Arid Environ* 69(2):352–358.
37. R Core Team (2013) *R: A Language and Environment for Statistical Computing* (R Foundation for Statistical Computing, Vienna, Austria), Version: 3.0.1.